Valorization of food waste leachates through anaerobic digestion.







1

2

3

4

5

6

22

23

24

30 31 32 33 38

Ioannis Kontodimos 1,*, Chrysovalantis Ketikidis 1 and Panagiotis Grammelis 1

- ¹ Centre for Research & Technology Hellas, Chemical Process and Energy Resources Institute,57001 Thessaloniki, Greece; kontodimos@certh.gr; ketikidis@certh.gr; grammelis@certh.gr;
- Correspondence: kontodimos@certh.gr

7 Abstract: According to European Union data, on average 173 kg per person of total food waste (organic waste) are produced annually, of which 92 kg per person come from households (organic 8 waste). Food waste is defined the waste from household, restaurants, canteens, food industries as 9 well as markets. The importance of food waste stretches from environmental pressures to economic 10 and social impacts. An Environmental technology for the biodegradation of food waste is anaerobic 11 digestion. Is a very attractive technique and combines waste treatment and renewable energy re-12 covery. This study investigates the characteristics of food waste leachates from composting buckets 13 and their valorization as substrate for anaerobic digestion process. 14

A complete characterization of different food waste leachates was conducted (pH, COD, VFAs, 15 heavy metals etc.). Food waste leachates proved to be an ideal feedstock for anaerobic digestion. In 16 this direction, batch tests were performed to evaluate the methane yield of food waste leachates 17 under different operating conditions. Three different SIR ratios were tested (0,5, 1,0 and 1,5). A SIR 18 equal to 0.5 proved to be the as the higher methane yield was achieved. The removal of COD under 19 all operating conditions was higher than 70% with the higher removal (85,18%) for an SIR equal to 20 1.5. 21

Keywords: food waste leachates ; Anaerobic digestion ; SIR ; VFA ; heavy metals ;

1. Introduction

Published: date

Publisher's Note: MDPI stays neutral with regard to jurisdictional claims in published maps and institutional affiliations.

Citation: Kontodimos, I .; Ketikidis, C. ; Grammelis, P. Valorization

of food waste leachates through an-

aerobic digestion. Appl. Sci. 2022, 12, x. https://doi.org/10.3390/xxxxx

Academic Editor: Received: date



Copyright: © 2022 by the authors. Submitted for possible open access publication under the terms and conditions of the Creative Commons Attribution (CC BY) license (https://creativecommons.org/license s/by/4.0/).

Anaerobic digestion is an environmentally favourable technology and the most 25 widespread good practice for the biodegradation of household waste. Anaerobic Diges-26 tion is a complex biochemical process in which organic material is decomposed by several 27 groups of microorganisms in the absence of oxygen while renewable energy such as bio-28 gas is generated. 29

This technique is very attractive, combines waste treatment and renewable energy recovery. In addition to these two benefits, anaerobic digestion also reduces the odor of waste material while the digestate is rich in nutrients that can be used as fertilizer after the process of fermentation.

It is important to mention that, for this technology, monitoring of significant factors 34 [1] is essential, such as: pH, Total Solids (TS), Volatile Solids (VS), Ammonium (NH4+), 35 Chemical Oxygen Demand (COD), Alkalinity, Volatile Fatty Acids (VFAs), Total Organic 36 Carbon (TOC) and Total Nitrogen (TN). 37

2. Materials and Methods

2.1 Sampling and pretreatment

Compost leachates were collected during the period May 2020 to May 2021. They are mainly consisting of fruits and vegetables, and their sampling procedure were from dif-42 ferent stages of the composting process. 43

Specifically, collected samples before their composting for analysis as raw materials 44 and leachate of compost samples in a pre-compost phase. The samples were collected 45 from specific sampling spots. Composting containers were placed in public markets for 46 the collecting of the food waste. In the period during October 2020 to May 2021 the sam-47 ples were leachates from mechanical composting plant and waste transfer stations. They 48 came from different cities of the Region of Western Macedonia, in Greece. Sample codes, 49 their origin and dates of sampling are presented in the following Table 1. For each sample, 50 1,5 L of compost leachates were collected and delivered to the laboratory of CERTH in 51 Ptolemais, within 24h. Depending on the test method, samples were filtered through membrane filter (glass fiber, 0,45 syringe filter etc.), acidified and centrifuged (HPLC anal-53 ysis). A portion of the filtrated liquid was freeze-dried, and the remaining was stored at 54 4°C before further analysis. 55

Table 1. Description of the Samples

Date of Sam- pling	Sample Name	Description of Sample			
1/6/2020	S1	Raw material from bucket of public market			
12/6/2020	S2	Raw material from bucket of public market			
22/6/2020	S3	Raw material from bucket of public market			
22/6/2020	S4	Raw material from bucket of public market			
3/11/2020	S5	Leachate from mechanical composting plants			
22/12/2020	S6	Leachate from mechanical composting plants			
17/3/2021	S7	Leachate from waste transfer station			
17/3/2021	S8	Leachate from waste transfer station			
28/4/2021	S9	Leachate from waste transfer station			
28/5/2021	S10	Leachate from waste transfer station			

2.2 Analytical methods

The measurements of TS, VS, COD, NH4⁺, Alkalinity, TOC and TN were carried out 58 according to APHA Standard Methods [2], The pH was measured using a digital pH-me-59 ter (Hanna, HI2260). The quantification of major and trace metals was carried out accord-60 ing to APHA Standard Methods and ISO 15586 and a Graphite Furnace Atomic Absorp-61 tion Spectrophotometer was used (Shimadzu GFA-EX7i AA-6300). For the quantifications 62 of TOC and TN, a TOC analyzer (Shimadzu, TOC-L) was used. 63

Finally, the quantification of VFAs was carried out by an HPLC. A portion of 100 ml 64 filtered sample was acidified with $30 \,\mu\text{L}$ of H₃PO₄ (HPLC grade) and centrifuged at 10.000 65 rpm for ten minutes Also, VFAs from anaerobic digestion process, were determined ac-66 cording to the Kapp method [3]. 67

2.3 High performance liquid chromatography

The used separation module Ecom, ECB2000, was equipped with pump and degasser 69 (Ecom, ECP2000), oven (Ecom, ECO2000), diode array detector (Ecom, ECDA) and cou-70 pled with a RP-C18 column (Fortis Technologies, 250X4,6mm, 5um). Spectra were ob-71 tained between 200 and 230 nm. Isocratic elution procedure applied to the mobile phase 72 (0,02mol/l KH₂PO₄/methanol) stable at 98:2 for 50 min. The mobile phase was acidified 73 with H₃PO₄ to reach pH 2,88. The flow rate was 0,6 ml/min at 35 °C and 20µl injection 74

2 of 6

39

40 41

52

56

57

volume [4]. Standard stock solution containing 100mg/l of the organic acids (Acetic, pro-75 pionic and butyric acids) prepared in ultrapure water. 76

2.4 Inoculum and Substrate

Anaerobic sludge was used as inoculum (microbial culture) for the biomethane potential test arrays and obtained from a commercial mesophilic anaerobic digester plant in the area of Eordea (Western Macedonia).

The substrate for the digestion process was the sample S9 (Table 1) because of the 81 C/N ratio (Table 2). Three ratios SIR (Substrate to Inoculum Ratio) were monitored for 82 biomethane potential, 1,5, 1,0 and 0,5. The calculation for the SIR ratios were determined 83 according to the VS of the substrate and the inoculum. . During start-up, flushing with N2 84 took place and all samples were incubated at mesophilic conditions (35 +/-2°C) throughout the experimental process. All batch tests were performed in triplicate.

2.5 Biomethane Potential Test

BMP tests are a technique to determine the methane potential and the biodegradabil-88 ity of any type of waste [5]. Batch experiments carried out using the Automated Methane 89 Potential Test System II (AMPTS II). Each of the AMPTS' bioreactors had 500 ml bottles 90 with 400ml working volume and 100 ml headspace, was equipped with an individual me-91 chanical stirrer and operated as a bench scale anaerobic glass bioreactor. The produced 92 biogas from each glass bioreactor passed through a 3 M NaOH solution which retained 93 CO2 and H2S. The upgraded biogas passed through a flow cell (one for each glass bioreac-94 tor) which measured gas productivity through water displacement. The digital impulse 95 was registered by a computer [6, 7]. The results of BMP test experiments are expressed as 96 normalized mL. 97

3. Results

The results of complete characterization of samples S1 to S10, are summarized below 99 in Tables 2 to 3. The significant variation of leachate characteristics could be attributed to 100 the impact of factors that affect quality and quantity, including waste composition, age of the waste and the composting technology used [8]. The leachate obtained was brown with 102 an unpleasant odor that could be attributed to the organic acids and volatile fatty acids 103 produced from composting food waste. Other volatile nitrogen and sulfur compounds 104 could also have contributed to this odor [9]. 105

The accumulated biomethane yields and the production flows of the three SIRs are 106 shown in Fig.1. Tables 4 and 5 shown the main characteristics of the inoculum and feed-107 stock. Figure 1 illustrates the accumulated Nml CH4/g VS (a) and the Flow rate (Nml/day) 108 of the three SIR ratios. Table 4 summarizes the composition of the SIR feedstock and inoc-109 ulum used in the batch test. Finally, table 5 depicts the main characteristics of each bench 110 scale bioreactor during the start-up phase and after the end of the batch experiment. 111

3 of 6

77

78 79

80

85 86

87

98



 Figure 1. (a)Accumulated Nml CH4/g VS added of SIR 0,5, 1,0 and 1,5. (b) Flow (Nml/day) of SIR
 113

 0,5, 1,0 and 1,5.
 114

Table 2. Complete characterization of compost leachate. VFAs represent the cumulated concentra-115tion of acetate, propionic and butyric acid.116

Parameter	S1	S2	S 3	S4	S 5	S6	S 7	S 8	S 9	S10
pН	4,29	4,65	4,52	3,84	4,02	4,82	4,35	4,91	4,48	4,05
TS (g/l)	118	24	69	114	36	18	33	26	41	60
VS (g/l)	98	14	45	100	28	14	25	17	31	48
COD (mg/l)	47202	49500	51800	38850	53924	16620	40154	25102	25601	31794
NH4+ (mg/l)	470	956	985	212	134	138	104	303	484	117
TOC (mg/l)	15580	14150	14140	34430	13930	7368	9990	19885	16660	3692
TN (mg/l)	2558	3435	3588	3973	579	267	314	511	771	273
VFAs (g/l)	30508	18349	16041	37046	41780	8847	12768	14874	21613	24946
C/N ratio	6,1	4,1	3,9	8,7	24,1	27,6	31,8	38,9	21,6	13,5

Table 3. Major and minor trace elements of compost leachate.

Parameter (mg/l)	S1	S2	S 3	S4	S 5	S6	S7	S 8	S9	S10
Na	6208	7210	7506	7005	3980	2075	233	577	552	1251
K	874	869	891	1329	145	280	1011	1320	1905	2779
Mg	337	227	226	532	370	166	90	316	168	895
Zn	1,4	2,1	2,1	4,3	4,0	1,4	12	27	18	16
Fe	203	335	340	164	281	357	1,4	0,5	31	32
Cu	0,3	0,3	0,5	0,6	0,6	0,2	0,6	0,3	1,1	1,6
Pb	nd	nd	nd	nd	nd	nd	nd	nd	nd	nd
Ni	nd	nd	nd	nd	nd	nd	1,5	1,0	1,2	2,5
Cr	nd	nd	nd	nd	nd	nd	1,2	3,4	3,3	2,6
Cd	nd	nd	nd	nd	nd	nd	nd	nd	nd	nd
Mn	nd	nd	nd	nd	nd	nd	1,2	1,2	3,8	3,4

Table 4. Composition of the SIR feedstocks and inoculum used in the batch experiments in bench118scale bioreactors.119

Parameter	Inoculum	FW (0,5)	FW (1,0)	FW (1,5)	
Alkalinity (mg/l	15000	3225	4150	5075	
CaCO3)	15000	3223	4150	5075	
VS (mg/l)	24000	2400	4800	7200	
COD (mg/l)	22820	8173	16047	23921	
NH4+ (mg/l)	972	67	83	103	
TOC (mg/l)	6623	952	1940	2930	

117

Table 5. Main characteristics of each bench scale bioreactor during the start-up phase and after the120end of the batch experiment. VFAs rep-resent the cumulated concentration of acetate, propionic and121butyric acid.122

Parameter	FW	(0,5)	FW	(1,0)	FW (1,5)	
	Initial	Final	Initial	Final	Initial	Final
	Concentration	Concentration	Concentration	Concentration	Concentration	Concentration
pН	6,41	7,95	6,03	7,79	5,80	7,88
VS (g/L)	19,54	2,65	31,6	4,24	47,38	4,71
VFAs (mg/L HACeq)	4371	1678	5175	1076	9366	948
NmL CH4/g VS	512	,00	511	,76	333	,02
COD (mg/L)	7138	1204	12451	3349	26880	3983
NH4 ⁺ (mg/L)	552	1053	818	1553	669	1355
Test Days	3.	2	6	0	6	0

4. Discussion

The pH values of samples ranged between 3,84 to 4,91. According to literature [10], 124 at the process of degradation of organic material, carbon dioxide and a low amount of 125 ammonia are produced and these two products have further resulted in the formation of 126 ammonium ions and carbonic acid. The carbonic acid dissociates to produce hydrogen 127 and bicarbonate ions, which influence the level of pH. 128

Solids (TS, VS) are influenced by the total amount of dissolved organic and inorganic129material. According to the literature a typical leachate ranged from 0,589-196 g/L[8]. The130present values in this study remain within this range, and the range of volatile solids is131between 13,5 g/L to 98 g/L.132

The range of TOC values is between 3.500 mg/L to 35.000 mg/L. The TOC content decreases during composting due to the microorganisms activity [11] and the further degradation of organic substances necessary for their metabolism.

Nitrogen is oxidized mainly to ammonium and to nitrite and subsequently, to nitrates when nitrification is achieved [10]. The values of ammonium range between 985 mg/L to 134 mg/L and Total Nitrogen between 3.973 mg/L to 267 mg/L.

The COD values include the oxygen demand created by biodegradable as well as 139 non-biodegradable substances. COD is highly variable and this is due to the food waste 140 composition and the climate characteristics [8], with reported values are varying between 141 16.620 and 53.924 mg/L. 142

Finally, the values of VFAs were present in high concentrations, that means the composting process is in an initial stage or it is a raw compost material and characterized as immature [12].

The batch experiments lasted 32 days for the SIR 0,5 and for 60 days SIR 1,0 and 1,5 146 respectively, until minimum or no biogas production was observed. Almost 2400 Nml 147 and 2456 Nml methane were produced from SIR 1,5 and 1,0 respectively, and 1326 Nml 148 methane from SIR 0,5. According to Figure 1a, the values of Nml methane correspond to 149 methane yields 333,02 Nml CH4/g VS added, 511,76 Nml CH4/g VS added and 512 Nml 150 CH₄/g VS added for SIR 1,5, 1,0 and 0,5 respectively. SIR 0,5 led to higher methane yield 151 in 32 days instead of the other two SIR in 60 days. The degradation of VS, VFAs and COD 152 for each SIR are shown in Table 5. 153

As illustrated in figure 1b, a high biogas flow rate was observed from the 1st day of 154 the three SIR batch experiments and continued until the 19st day. Furthermore, the flow 155 reduction for SIR 0,5 ceased at the 32nd day. SIR 1,0 displays a high flow rate until the 30th 156 day. After the 30th day a continuous reduction of biogas flow rate was observed until it 157 stops at 60th day. Finally, SIR 1,5, after the 30th day displays a continuous increase of biogas 158 flow rate until the cease of the batch experiments test at the 60th day. 159

123

133

134

135

136

137

138

In this study the changes of food waste compost leachate were monitored with regard 161 to the seasonality, the composting time and the biomethane yield in three different SIR. 162 The compost leachate showed a high organic load that means an ideal substrate for com-163 posting or anaerobic digestion, but simultaneously showed high values of Volatile Fatty 164 Acids, that means the compost is in the initial stages of the composting process and con-165 sidered as immatured. Regarding the values of COD and total nitrogen, in this stage of 166 the process, they are in high concentrations thus recommended to use the leachate in low 167 application rates or after dilution. 168

Finally, food waste compost leachates could be characterized as an ideal substrate for 169 anaerobic digestion. The three different food waste SIR ratios in the bench scale experiment, produced in sufficient quantity (expressed as Nml CH₄ /g VS added) biomethane, 171 however the SIR 0,5 produced higher biomethane yield at the half days of the procedure. 172 The degradation of the COD rates of the SIR 0,5, 1,0 and 1,5 were 83,13%, 73,11% and 173 85,18% respectively. 174

Author Contributions: I. Kontodimos: Methodology, Validation, Investigation, Writing-review & editing. C. Ketikidis: Writing, review & editing, Project administration, P. Grammelis: Writing-review & editing, Supervision.

Acknowledgements:The main parts of this work were carried out within the Less Waste II project178(https://ii.less-waste.eu/) that was co-funded by the European Union and national funds of the179participating countries within the Interreg IPA CBC Programme Greece – Albania 2014-2020 (Subsidy Contract No A2 – 1.1 - 5)180

Conflicts of Interest: "The authors declare no conflict of interest."

References

1.	B.Wang, Factors that Influence the Biochemical Methane Potential (BMP) Test, Steps towards the Standardisation of BMP Test,	184
	Doctoral Dissertation, Faculty of Engineering, Lund University, Sweden 2016	185
2.	APHA/AWWA/WEF. Standard Methods for the Examination of Water and Wastewater, Stand. Methods. (2012) 541.	186
	https://doi.org/ISBN 9780875532356	187
3.	V.T. Mota, F.S. Santos, T.A. Araujo, M.C.S. Amaral. Evaluation of titrations methods for volatile fatty acids measurement: effect	188
	of the bicarbonate interference and feasibility for the monitoring of anaerobic reactors. Water Practice & Technology Vol 10 No	189
	3 (2015) doi: 10.2166/wpt.2015.056	190
4.	A.C.M.de Sena Aquino, M.S. Azevedo, D.H.B. Ribeiro, A. C.O. Costa, E.R. Amante. Validation of HPLC and CE methods for	191

- 4. A.C.M.de Sena Aquino, M.S. Azevedo, D.H.B. Ribeiro, A. C.O. Costa, E.R. Amante. Validation of HPLC and CE methods for 191 determination of organic acids in sour cassava starch wastewater. Food Chemistry (2014) http://dx.doi.org/10.1016/j.foodchem.2014.09.142 193
- J.Filer, Huihuang H. Ding, S. Chang. Biomechemical Methane Potential (BMP) Assay Method for Anaerobic Digestion Research.
 Water (2019). doi:10.3390/w11050921
 195
- 6. B.Hulsemann, L.Zhou, W. Merkle, J.Hassa, J. Muller, H. Oechesner. Biomethane Potential Test: Influence of Inoculum and Digestion System. Appl.Sci.(2020) doi:103390/app10072589
- G.Lytras, E. Koutroumanou, G. Lyberatos. Anaerobic co-digestion of condensate produced from drying of Household Food 198 Waste and Waste Activated Sludge. Journal of Environmental Chemical Engineering 8 (2020) 103947 199 https://doi.org/10.1016/j.jece.2020.103947 200
- S.Q. Aziz, H.A. Aziz, M.S. Yusoff, M.J.K. Bashir, M. Umar, Leachate characterization in semi-aerobic and anaerobic sanitary landfills: A comparative study, Journal of Environmental Management 91 (2010) 2608-2614, doi:10.1016/j.jenvman.2010.07.042
 202
- 9. P.M.Sall, H. Antoun, F.P.Chalifour, C.J. Beauchamp. Potential use of leachate from composted fruit and vegetable waste as fertilizer for corn. Cogent Food & Agriculture (2019), 5: 1580180. https://doi.org/10.1080/23311932.2019.1580180
- 10.
 B.P. Naveen, D.M. Mahapatra, T.G. Sitharam, P.V. Sivapullaiah, T.V. Ramachandra, Physico-chemical and biological characterization of urban municipal landfill leachate. Environmental Pollution 220 (2017) 1-12 http://dx.doi.org/10.1016/j.envpol.2016.09.002
 205
- 11. Y.A.Y. Abdellah,Z.J. Shi,Y.S. Luo,W.T. Hou,X. Yang,R.L. Wang. Effects of different additives and aerobic composting factors on heavy metal bioavailability reduction and compost parameters: A meta-analysis. Environmental Pollution 307 (2022) 119549
 208

 https://doi.org/10.1016/j.envpol.2022.119549
 210
- P. Prochazka, Acute and subchronic phytotoxicity of volatile fatty acids, Master of Science, Environmental Science, Faculty of Mathematics and Science, University of Jyvaskyla, Finland (2008)

182

175

176

177

183

196

197

203