Carbon dioxide and methane emissions during composting
and vermicomposting of sewage sludge under the effect of
different proportions of straw pellets

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# 5 Bayu Dume<sup>a\*</sup>, Ales Hanc<sup>a</sup>, Pavel Svehla<sup>a</sup>, Abraham Chane<sup>a</sup>, Abebe 6 Nigussie<sup>b</sup>

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8 <sup>a</sup>Czech University of Life Sciences, Faculty of Agrobiology, Food, and Natural

9 Resources, Department of Agro-Environmental Chemistry and Plant Nutrition,

10 Kamycka 129, Prague 16500, Czech Republic. \*<u>dumebayu@gmail.com</u>

11 <sup>b</sup>Jimma University, College of Agriculture, 307, Jimma, Ethiopia

12 Abstract: The aim of this study was to evaluate the carbon dioxide (CO<sub>2</sub>) and methane (CH<sub>4</sub>) emissions during composting and vermicomposting of 13 sewage sludge under the effect of different proportions of straw pellets. 14 Four treatments, including a control with three replicates, were designed 15 to mix the initial sewage sludge with varying rates of pelletized wheat 16 straw (0, 25%, 50% and 75% (w/w)). Over a 60-day period, 17 vermicomposting with Eisenia Andrei treatments and composting were 18 carried out. The results indicated that both composting 19 and vermicomposting produce a significant (p<0.001) amount of CO<sub>2</sub> and CH<sub>4</sub> 20 emissions from all treatments. Vermicomposting significantly reduced 21 CH<sub>4</sub> emissions by 18%, 34%, and 38% and increased CO<sub>2</sub> emissions by 22 75%, 64%, and 89% for the treatments containing 25%, 50%, and 75% straw 23 pellets respectively, compared to composting. However, CO<sub>2</sub> emissions 24 decreased and CH<sub>4</sub> emissions increased during composting compared to 25 vermicomposting. As a result of this finding, both composting and 26 vermicomposting processes are recommended as an additive of pelletized 27 wheat straw, depending on the target gas to be reduced. 28

29 Keywords: Thermophilic; earthworms; biosolids; greenhouse gases;

30 composting

# 31 **1. Introduction**

Sewage sludge is the residual, semi-solid material that is produced as a by-product during the process of biological wastewater treatment or municipal waste-water. The large amounts produced in the recent decades represent an increasing, and improper disposal or management has resulted in a serious environmental pollution due to the putrescible nature of sewage sludge and waste management challenges [1]. The improper

management of sewage sludge will cause secondary pollution such as 1 pathogenic microbes, organic micro-pollutants, and toxic heavy metals. 2 Therefore, sustainable and eco-friendly sewage sludge management is 3 urgently required [2]. According to He et al. [3], currently, the annual 4 production of sewage sludge in the European Union reaches over 10.96 5 million tons per year and 40 million tons in China [4]. This amount is 6 increasing due to expedited urbanization and the increasing capacity of 7 municipal wastewater treatment facilities [5]. 8

Composting and vermicomposting are effective techniques and low 9 cost methods to manage and reuse sewage sludge due to its safe and 10 stable products that could be used as an organic fertilizer or soil 11 conditioner for farming [6]. However, harmful gases, such as ammonia 12 (NH<sub>3</sub>), nitrous oxide (N<sub>2</sub>O), and methane (CH<sub>4</sub>), are emitted due to the 13 mismanagement of sewage sludge. CH4 and CO2 are two of the most 14 important greenhouse gases in the atmosphere. CH<sub>4</sub> is radiatively stronger 15 than CO<sub>2</sub> on a mass basis and it is reported that the current global 16 warming potential of CH<sub>4</sub> is 25 times higher than that of CO<sub>2</sub> over a 100 17 [7]. Most previous studies 18 vear period on composting and vermicomposting have focused on the feasibility of different organic 19 wastes, the factors affecting the growth and reproduction rate of 20 earthworms, as well as the quality of compost and vermicomposts [8]. 21 However, little is known about the emissions of CO2 and CH4 during 22 composting and vermicomposting of sewage sludge. Therefore, the aim of 23 this study was to evaluate the carbon dioxide (CO<sub>2</sub>) and methane (CH<sub>4</sub>) 24 emissions during composting and vermicomposting of sewage sludge 25 under the effect of different proportions of straw pellets. 26

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## 2. Materials and methods

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#### 29 2.1. Raw materials

The experiment was carried out at the experimental station of the 30 Faculty of Agrobiology, Food and Natural Resources, Czech University of 31 Life Science, Prague, in Cerveny Ujezd. The sewage sludge used in the 32 experiments was collected from the waste-water treatment plant in the 33 Czech Republic. Dried pelletized wheat straw was provided by Granofyt 34 Ltd Company with a diameter of 10mm. The selected chemical properties 35 of sewage sludge and pelletized wheat straw are listed in (Table 1) and for 36 the treatments at the initial days (day-0) in (Table 2). Eisenia Andrei was used in 37 this study for vermicomposting. 38

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Parameters	Sewage Sludge(SS)	Pelletized wheat Straw(PWS)
pH-H <sub>2</sub> O	6.99±0.03	8.30±0.52
EC(mS/cm)	0.617±0.11	0.680±0.07
TOC (%)	32.95±0.26	42.6±0.36
TN (%)	5.36±0.03	0.8±0.12
C·N	6 15+0 04	53 2+7 60

Table 1. Selected chemical properties of initial materials

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4 Table 2. Selected chemical properties of treatments at the initial (day-0)

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Treatments	pH-H <sub>2</sub> O	EC(mS/cm)	TOC (%)	TN (%)	C: N
T1	6.99±0.03	0.617±0.11	32.9±0.26	5.36±0.03	6.14±0.04
T2	7.32±0.11	0.633±0.08	35.36±0.23	$1.98 \pm 0.21$	18.03±1.92
T3	7.64±0.25	0.649±0.06	37.77±0.24	$1.34 \pm 0.07$	28.17±1.43
T4	7.97±0.38	$0.664 \pm 0.05$	40.18±0.29	$1.05 \pm 0.05$	38.36±2.03

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# Values indicate mean ± standard deviation (n =3)

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# 2.2.Experimental design

The experiment included four treatments with three replications: (T1) 8 100% sewage sludge (control), (T2) 75% sewage sludge (SS) + 25% 9 pelletized wheat straw (PWS), (T3) 50% sewage sludge (SS) + 50% 10 pelletized wheat straw (PWS), (T4) 25% sewage sludge (SS) + 75% 11 pelletized wheat straw (PWS) (w/w). The pelletized wheat straw was 12 applied on a wet weight basis. In all the treatments, the substrate was 13 homogenized and transferred to fermenter barrels for 60 days for 14 composting and also the same treatments were transferred to worm-bins 15 for vermicomposting. Each worm-bin received 377(57.4g) pieces of adult 16 earthworms (Eisenia andrei). The moisture level of the material was 17 maintained at about 70-80% of wet mass throughout the vermicomposting 18 stage by spraying the surface with water at two-day intervals. 19

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# 2.3. Measurements of carbon dioxide(CO<sub>2</sub>) and methane(CH<sub>4</sub>) during composting and vermicomposting

Concentrations of CO<sub>2</sub> and CH<sub>4</sub> during both composting and vermicomposting were measured by a closed chamber technique. A tightfitting lid with two ports for headspace gas sampling and air temperature measurement was used to connect one side tip of plastic tube to closed barrels for composting and a worm bin for vermicomposting, and the other side tip of plastic tube was connected with instruments during data recording. Measurements were done twice per day within 12hour
 intervals for 60 days by using the Gasko Infrared Gas Analyzer [9].

To calculate the cumulative CO<sub>2</sub>, and CH<sub>4</sub> emissions, we summed daily values to get the total cumulative gas emissions during the whole experimental period [9].

$$A_{t(ab)} = \frac{(t_b - t_a).(F_{ta} + F_{tb})}{2}$$
(1)

6 Where A<sub>t(ab)</sub> is the cumulative emission between the measurement days
7 (between ta and t<sub>b</sub>), ta and t<sub>b</sub> are the measurement dates, and F<sub>ta</sub> and F<sub>tb</sub> are
8 the gas fluxes on the two measurement dates. Therefore, the total
9 cumulative emissions were calculated as the sum of cumulative emissions
10 on each day using Equation (2):

$$Total cumulative emission = \sum A_{t(ab)}$$
(2)

## 11 2.4.Analysis of total carbon (TOC), total nitrogen (TN), pH, and EC

The samples were taken for determination of TOC, TN, pH and EC, using standard methods. pH and electrical conductivity (EC) were measured in distilled water at 1:5(w/v). The values of total carbon (TOC) and total nitrogen (TN) were acquired with an elemental analyzer (Elemental Vario EL, German).

#### 17 2.5. Statistical analyses

The statistical analyses were carried out using the R version 4.0.2 statis tical package. ANOVA was used to test the significant sources of variation , and the following Tukey HSD test was used to compare the treatment me ans if the factors' effect was significant at P < 0.05. Two-way analysis of va riance (ANOVA) was performed to analyse the significant differences bet ween treatment and composting process methods.

#### 24 **3. Results and Discussions**

### 25 3.1.Temperature during composting

The temperature in each treatment reached its maximum during the composting process, with the significant differences between treatments (Figure 1). Variations in the temperature were the result of mixing with different percentages of pelletized wheat straw. The temperature of two treatments (T3 and T4) rapidly reached the thermophilic stage (>50°C) on

days 3 and 2 respectively. T4 reached the maximum thermophilic phase of 1 65.5°C in four days and 57.4°C for T3. The thermophilic phase lasted for 14 2 days in T4, and 10 days in T3. The maximum temperature for the 3 remaining treatments was 37.6°C for T2 and 29.55°C for T1, temperatures 4 gradually declining until the end of the experiments. Thus, the addition of 5 pelletized wheat straw resulted in more intensive decomposition in the 6 thermophilic phase, but in the cooling phase, the degradation process 7 resulted in less heat in these mixtures due to the depletion of easily 8 degradable organic compounds [10]). T1 (control) and T2 (25%PWS) 9 delayed reaching the thermophilic stage and had no thermophilic phase at 10 all, and the maximum temperature was 37.6°C for T2 and 29.55°C for 11 control and lasted to maturity within the mesophilic temperatures. This 12 might be due to the high moisture in these treatments. 13





16 *3.2.pH and EC* 

The pH of final compost and vermicompost for all treatments are 17 showed in (Table 3). The proportions of pelletized wheat straw in the 18 mixtures resulted in lower pH values during vermicomposting and this is 19 probably due to the high content of organic acids (e.g. Succinic and Maleic 20 acid) and directly proportional to the amount of straw in the 21 22 treatments[11]. The pH of the compost (T1, T2, T3, and T4) was higher than vermicompost. However, the pH in vermicompost has decreased 23 significantly (p<0.05). The similar pH behavior during vermicomposting of 24 sewage sludge, crop straw, municipal solid waste, and livestock manure 25 was also reported by other researchers [11]. The release of low molecular 26 weight organic acids from organic decomposition and the increase in 27 nitrification could decrease the pH during vermicomposting [12]. A 28

decrease in pH during vermicomposting of different feeding materials has
 been reported [13, 14]. The lower pH of vermicompost might indicate that
 a more intense decomposition reaction is undergone during
 vermicomposting than in composting.

5 Table 3. Selected chemical properties of end product compost and6 vermicompost

Processes	Treatments	pH-H <sub>2</sub> O	EC(mS/cm)	TOC (%)	TN (%)	C:N
С	T1	8.43±0.12	1.90±0.17	29.52±0.73	$4.55 \pm 0.14$	6.50±0.04
	T2	8.32±0.09	1.43±0.09	32.43±0.79	3.69±0.03	8.84±0.32
	T3	8.35±0.08	$1.94 \pm 0.14$	34.45±1.53	3.27±0.05	10.57±0.65
	T4	8.01±0.06	0.80±0.06	37.95±0.02	2.76±0.15	13.88±0.80
VC	T1	6.66±1.16	$0.644 \pm 0.04$	28.43±0.32	4.22±0.20	6.77±0.26
	T2	6.47±1.5	1.186±0.22	31.96±0.89	$3.58 \pm 0.04$	8.94±0.35
	Т3	$6.50\pm0.14$	0.802±0.39	34.38±1.13	2.95±0.15	11.72±0.93
	T4	6.65±0.31	1.21±0.12	35.32±0.37	3.08±0.06	12.15±0.32

7 C=composting, VC= vermicomposting, values indicate mean ± standard deviation (n =3),

The EC value was higher in compost than in vermicompost made 8 from the same raw material and treatments (Table 3). The EC gradually 9 increased in all of the treatments, which could be explained by the release 10 of bonded elements during earthworm digestion [15, 16], and the release 11 of minerals during the decomposition of organic matter in the form of 12 cations in the vermicompost [17]. The final EC was within the 13 recommended limit of 2dS/m [18] for all the treatments, which indicates 14 an ideal vermicompost/compost for application to plants. The increased 15 EC during the period of vermicomposting processes is in consistency with 16 that of earlier workers [19, 20], which was probably due to the degradation 17 of organic matter releasing minerals such as exchangeable Ca, Mg, K, and 18 P in the available forms, that is, in the form of cations in the vermicompost 19 and compost [17]. 20

# 3.3. Emissions of CO<sub>2</sub> and CH<sub>4</sub> during composting and vermicomposting 3.3.1. Carbon dioxide (CO<sub>2</sub>)

The CO<sub>2</sub> emissions increased at the beginning of composting and vermicomposting (Figure 2a, c) because of rapid degradation of easily degradable organic matter and thereafter gradually decreased until the end of composting/vermicomposting. This finding reveals the findings reported by Awasthi et al. [21] and Meng et al. [6] during the composting of the sewage sludge. During the first 13 days, the CO<sub>2</sub> emissions in control (T1) were higher than the other treatments (T2, T3, and T4) during

composting. But, the CO<sub>2</sub> emissions in this treatment T1 is lower during 1 vermicomposting. This result was possible because the earthworms 2 inhibited microbial activity and reduced the readily available OM 3 [22].Significant differences were found between the four treatments and 4 the composting/vermicomposting process (P <0.001). These findings 5 indicate that pelletized wheat straw may be lost in the inhibition after the 6 thermophilic stage, most likely due to self-degradation at high 7 temperature [23]. The temperature and pH of T1, T2, T3, and T4 also 8 9 support this conclusion. A sharp drop in CO<sub>2</sub> emissions on day 14 and a small peak on day 20 appeared in all treatments (Figure 2a). 10





This observation could be attributed to the anaerobic environment caused by the strong degradation of OM during the first 14 days. The subsequent turn on day 10 destroyed the anaerobic conditions. Similar results were also reported in previous studies [21] for sewage sludge composting.

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#### 3.3.2. *Methane* (*CH*<sub>4</sub>)

2 CH<sub>4</sub> from all treatments during the composting and vermicomposting processes is displayed in (Figure 2b, d). The results of this study showed 3 that CH<sub>4</sub> concentrations for all treatments peaked relatively early in both 4 composting and vermicomposting processes within 1-3 weeks, after which 5 emission rates gradually declined until the end of the experiment. 6 Therefore, it could also be assumed that the CH<sub>4</sub> emissions should also be 7 the highest during the start of the process. Several researchers reported 8 similar findings, namely that the highest levels of CH<sub>4</sub> emissions occurred 9 at the start of the composting and vermicomposting processes [24]. CH<sub>4</sub>, a 10 major GHG generated during composting and vermicomposting, is a 11 significant contributor to global warming. The production of CH4 is 12 attributed to methanogen deoxidization of CO<sub>2</sub>/H<sub>2</sub> and acetic acid under 13 low oxygen conditions [25]. 14



Figure 3. Total cumulative emissions of CO<sub>2</sub>-C (a), CH<sub>4</sub>-C (b) after 60 days of 17 composting, CO<sub>2</sub>-C(c), CH4-C (d) during vermicomposting. Bars indicate the 18 standard error of the means (n=3). Different letters indicate significant differences 19 among the treatments (*p*<0.05) 20

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Thereafter, as an organic matter (OM) decomposed and oxygen was 1 replenished through turning, the CH4 emissions of all treatments declined 2 sharply and remained at a low level during the maturation phase of 3 composting and vermicomposting. The pattern of CH4 emissions observed 4 resembles the patterns reported by Ma et al. [26] and Wang et al. [27]. 5 Since microorganisms can rapidly degrade organics in the thermophilic 6 phase, a dramatic reduction in O<sub>2</sub> levels can be observed in the compost 7 [28]. In all treatments, the emission of CH<sub>4</sub> is higher during composting 8 than vermicomposting and the higher results are measured in the control 9 area. 10

Total cumulative CO<sub>2</sub> differed by composting method (P<0.001), as did their interaction (P<0.001) (Figure 3). Vermicomposting increased total cumulative CO<sub>2</sub> emissions when compared with thermophilic composting. Composting had an effect on total cumulative CH<sub>4</sub> emissions (P<0.001). Vermicomposting decreased CH<sub>4</sub> emissions by 74.5% from a high proportion of pelletize wheat straw T4 compared with thermophilic composting.

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#### 3.4.Total Organic Carbon(TOC), Total Nitrogen(TN) and C: N ratio

The TOC, TN and C: N ratio content for all treatments is presented in 19 (Table 3). It can be seen that the TOC and C: N contents decreased in both 20 compost and vermicompost as compared with initial treatments. 21 22 However, the TN content increased in both compost and vermicompost. The increase in TN content was caused by the loss of ammonia 23 volatilization at relatively high temperatures and a pH that was not 24 suitable for nitrification and denitrification [29]. Zhang et al. [30] 25 attributed the increase in TN during vermicomposting of sludge and the 26 increase was due to the activity of worms. C: N ratio for all treatments 27 decreased with both composting and vermicomposting processes. The C: 28 N ratio indicates the maturity of compost/vermicompost since it reflects 29 stabilization and mineralization rates during vermicomposting [31]. Our 30 results are corroborated by previous studies by [32] who reported up to 31 50.86% and 48.8% reduction in C: N ratio during vermicomposting of cow 32 dung, and cow dung with vegetable waste, respectively. The final C: N 33 ratio recorded for all the treatments was within the recommended value 34 for soil applications <20 [33]. 35

## 1 4. Conclusions

The composting and vermicomposting processes of sewage sludge 2 emitted a considerable amount of CH<sub>4</sub> and CO<sub>2</sub>, the main environmental 3 threat to global climate change. The highest values were at the beginning 4 of the experiment and gradually decreased. The emission of CH<sub>4</sub> and CO<sub>2</sub> 5 during composting and vermicomposting is linked to the fate of C present 6 7 in the waste substrate. Vermicomposting reduces CH4 emissions and accelerates the decomposition process. The addition of different 8 proportions of pelletized wheat straw increases CO<sub>2</sub> and CH<sub>4</sub> emissions 9 during composting. Vermicomposting increases CO<sub>2</sub> emissions, implying 10 that vermicompost is at a more advanced stage of decomposition than 11 thermophilic compost. From this finding, as an additive of pelletized 12 wheat straw, both composting and vermicomposting processes are 13 recommended depending on the target gas to be reduced. 14

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