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Temporal dynamics of vegetation indices for fires of various severities in southern Siberia

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Proceedings

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Abstract: Wildfire is a critical environmental disturbance affecting forest dynamics, succession, and 10 the carbon cycle in Siberian forests. In recent decades forests of southern and central Siberia expe-11 rienced an increase in fire-disturbed area. The main goal of this study was to assess the degree of 12 fire disturbance in the southern regions of central Siberia, as well as the dynamics of post-fire 13 changes for fires of different intensity. Remote sensing data from MODIS and VIIRS sensors were 14 used to estimate burned area, fire radiative power (FRP) and post-fire dynamics using Normalized 15 Burn Ratio (NBR) and Normalized Difference Index Vegetation (NDVI). Mean annual forest 16 burned area between 2001 and 2021 in the region was about 250 thousand ha per year with the 17 largest burned areas observed in mixed and larch-dominant forests. Fires detected in the 18 dark-needle coniferous (DNC) and larch-dominant forests were found to have higher (by about 19 25%) fire radiative power comparing to fires in pine-dominant and mixed forests. The analysis of 20 FRP together with NBR showed a significant correlation ($R^2 = 0.46$; p < 0.05) between these varia-21 bles, indicating that fires with higher intensity generally result in higher degree of fire disturbance. 22 Evaluation of the post-fire dynamics showed that NBR is more sensitive to fire-related disturbances 23 comparing to NDVI and requires more than 16 years to return to pre-fire values. At the same time, 24 in case of the NDVI the difference between disturbed and background areas was less than 1σ after 25 11 years since fire. The study was supported by the Russian Science Foundation and the Govern-26 ment of Republic of Khakassia (grant #22-17-20012, https://rscf.ru/en/project/22-17-20012/). 27

Keywords: Siberia; wildfires; satellite data; fire radiative power; vegetation index; NBR; NDVI

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1. Introduction

Wildfire is one of the dominant disturbances influencing vegetation dynamics, bio-31diversity and carbon cycling in boreal forests of Russia [1, 2]. Burned area across Russia32can vary greatly between years depending on weather conditions, with a mean annual33forested burned area of 5–7 million ha [1]. In particular, forests of southern and central34Siberia experienced an increase in area burned by wildfires in recent decades [3, 4].35

Assessments of burn severity can be considered as important factors for better un-36 derstanding of wildfire effect of forest ecosystems. For instance, fires in pine forests of 37 southern Siberia usually do not lead to significant tree mortality [5], while fires in 38 dark-needle coniferous stands result in complete tree mortality [6]. Thus, the assessment 39 of the regional peculiarities of the degree of forest disturbance caused by fires is still an 40 urgent task. Time period required for vegetation indices to recover after wildfire de-41 pends on forest type, wildfire severity, climate and the initial post-fire density of tree 42 seedlings [7, 8]. A better understanding of postfire dynamics will contribute to predicting 43 the effects of the increasing number of wildfires observed under climate change. 44

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Copyright: © 2022 by the authors. Submitted for possible open access publication under the terms and conditions of the Creative Commons Attribution (CC BY) license (https://creativecommons.org/license s/by/4.0/). The aim of the study was to estimate the fire disturbance of forests in the southern 45 regions of the Central Siberia between 2001 and 2021 including the following objectives: 46 (1) to estimate the proportion of fire-disturbed forest stands for several forest types predominant in the region; (2) to analyze the relationship between fire radiative power and 48 forest disturbance degree estimated using dNBR index; (3) evaluate the postfire dynamics of burned areas comparing to unburned areas. 50

2. Materials and Methods

2.1 Study area

The study area covers southern regions of the Central Siberia between $50 - 58^{\circ}$ N and 54 86 – 99°E. The area of the study region was about 7.5×10^5 km² (Figure 1). According to 55 the vegetation map developed by Space Research Institute and available at 56 http://pro-vega.ru/maps/, the dominant tree species include dark-needle coniferous 57 (DNC) forests mainly represented by cedar (Pinus sibirica) and fir (Abies sibirica) (25% of 58 the study area) with a smaller proportion of larch (Larix sibirica) (16%) and pine (Pinus 59 sylvestris) (5%). A significant part of the forest area (12%) is occupied by mixed forests 60 with a predominance of deciduous species (Betula spp., Populus tremula) [9]. 61



Figure 1. Study area. Land cover types are shown in colors.

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2.2 Data

Data products generated from the MODIS data for 2001-2021 were used to locate 66 burned areas and evaluate the degree of pyrogenic disturbance. Burned areas were de-67 lineated using the MODIS burned area product (MCD64A1) with the spatial resolution of 68 500 m [10]. Surface reflectance product (MOD09A1) with the spatial resolution of 500 m 69 [11] was used to estimate the degree of vegetation disturbance by fires. To estimate fire 70 radiative power (FRP) the MODIS thermal anomalies and fire product (MOD14A1 with 71 spatial resolution of 1000 m) was used [12]. Data products were downloaded using the 72 LAADS service (Level-1 and Atmosphere Archive & Distribution System, 73 https://ladsweb.modaps.eosdis.nasa.gov). 74

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2.3 *Methods* Data processing included an assessment of fire radiative power and the degree of pyrogenic disturbance of forests for each fire pixel.

Prevailing forest types within the study region were determined using the vegeta-

tion map developed by Space Research Institute and available through the VEGA service

The delta normalized burn ratio (dNBR) calculated from the MODIS surface reflectance product was used to estimate the degree of fire-caused vegetation disturbance [13, 14]. This index was calculated as the difference between pre-fire (the year preceding the fire) and post-fire (the year following the fire) normalized burn ratio (NBR) values. According to the previous classification [14], fire-disturbed areas with a dNBR value of 0.44 and higher can be characterized as highly disturbed. 87

Fire dates and locations were obtained from the MODIS burned area product. For 88 each year between 2001 and 2021 GIS layers of burned area were generated resulting in 89 21 burned area raster layers. For this study only fires on forest lands were considered, 90 while fires on non-forest lands (steppe, agricultural lands) were excluded. Using the 91 thermal anomalies product FRP values were obtained for each fire pixel. If several FRP 92 values corresponded to one fire pixel the maximum value was used. The entire FRP 93 range was divided into 50 MW/km² intervals and for each interval, the number of fire 94 pixels was calculated, which characterizes the frequency of fire occurrence with a given 95 FRP value, as well as its mean value. 96

For each season between 2001 and 2021 time series of spectral indices (NBR and 97 NDVI) were created for the period from mid-June to the end of August (161-233 days of 98 the year) excluding low quality data. For these time series the mean values for each fire 99 seasons were calculated. Thus, for each fire pixel there were 21 values of the NDVI and 100 NBR. Spectral indices for fire-disturbed areas were analyzed in comparison with undis-101 turbed (background) sites in similar conditions. Background values were calculated for 102 areas 100x100 pixels in size (~50x50 km) around burned areas. For these background ar-103 eas mean values and standard deviations were calculated for both spectral indices. A 104 pixel-based analysis of deviations from background values was performed. The devia-105 tion of the spectral indices from the background was calculated using Z-scores from the 106 ratio 107

$$Z = \frac{\sqrt{(V - V_b)^2}}{\sigma_b}$$
¹⁰⁸

where V and V_b – represent the post-fire and background values of spectral indices, 109 and σ_b – standard deviation of the background value. 110

3. Results and Discussion

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(http://pro-vega.ru/maps/) [5].

Annual burned areas in the region between 2001 and 2021 were highly variable de-112 pending on weather conditions ranging between 31 and 537 thousand hectares with the 113 decreasing trend in burned area ($R^2 = 0.29$; p < 0.05). Mean annual burned area was 114 250.2±153.7 thousand ha (mean±SD). The highest disturbance rate (11.4%) calculated as 115 ratio of the total burned area to the total forest area was observed in mixed forests, 116 meaning that 11.4% of mixed forest experienced fire disturbance between 2001 and 2021. 117 The lowest (3.1%) was observed in dark-needle coniferous stands (Table 1). The highest 118proportions of severely burned area was observed in the larch-dominant (20.8%) and 119 DNC forests (19.1%) while in the pine-dominant stands and mixed forests these were 120 about two times lower (10.9% and 9.8%, respectively) (Table 1). 121

Table 1. Fire disturbance (2001 – 2021) for main forest types.

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Dominant tree species	Burned area as fraction of total forest area, %	Severely burned area as fraction of total burned area, %
Larch	7.2	20.8
Mixed	11.4	9.8
Pine	6.0	10.9
DNC	3.1	19.1

The distribution of the fire radiative power was approximated by a power law (R^2 = 123 0.92, p < 0.05) (Figure 2a) that is consistent with previous results obtained for the boreal 124 forests of Eurasia and North America [15, 16]. Mean FRP for the study area was 37.4±26.4 125 MW/km², the lowest FRP values were observed for pine and deciduous stands (33.8±34.5 126 MW/km² and 30.1±29.3 MW/km², respectively) (Figure 2a). In the larch and DNC FRP 127 values were 25-30% higher making 46.6±42.7 MW/km² and 42.8±44.1 MW/km², respec-128 tively. These results were in good correspondence with the previously obtained estimates 129 for the central and northern regions of Siberia [16] and were approximately 20% higher 130 than the estimates for the Altai-Sayan region [16]. The Mann-Whitney U-test showed that 131 differences in FRP between forests with different dominant species should are significant 132 at the level of 0.05 (Figure 2a). 133



Figure 2. (a) Frequency distribution of fire pixels depending on their FRP. The numbers indicate 134 different dominant tree stands: 1 - larch forests (red circles); 2 - deciduous forests (blue diamonds); 135 3 – pine forests (purple squares); 4 – DNC forests (cyan triangles). (b) dNBR depending on the FRP. 136 Color scheme is the same as for the left panel. Error bars correspond to one standard deviation. 137 Each point shows the mean dNBR value for the corresponding FRP interval. Points are shown only 138 for FRP intervals with 50 fire pixels or more. 139

FRP values were also compared with the corresponding dNBR values. For fire pixels 140 from each 50 MW/km² interval the mean FRP values were calculated, as well as dNBR 141 means and standard deviations. Fire pixels within each FRP range were characterized by 142 significant dNBR variations (standard deviation in Figure 2b). The relationship between 143 FRP and mean dNBR values within these intervals can be approximated by a logarithmic 144law ($R^2 = 0.46$; p < 0.05) (Figure 2b). For example, for larch stands, an increase in the FRP 145 from 50 to 750 MW/km² resulted in increase in the dNBR by about 43% (Figure 2b), 146 however, further FRP increase resulted in much smaller increase in dNBR. FRP change 147 from 750 to 1750 MW/km² led to an increase in dNBR by only about 8%. Is also should be 148 noted that in severely burned areas (dNBR > 0.44) mean FRP value was 72–116% higher 149 comparing to moderately burned areas. 150

Post-fire dynamics of NBR and NDVI showed significant differences between areas 151 with high and moderate disturbance degree (estimated using dNBR). For instance, for 152

NBR index in the first post-fire year Z-value was almost two times higher for severely153disturbed areas comparing to moderately disturbed areas (Figure 3). At the same time154this initial anomaly was smaller in case of NDVI – severely disturbed areas had 1.5–2155times higher anomaly values compared to moderately disturbed areas.156



Figure 3. Post-fire dynamics for: (a) NBR and (b) NDVI. Solid lines correspond to severely dis-157turbed areas while dashed lines show moderately disturbed areas. Error bars correspond to one158standard deviation.159

In the case of moderately disturbed areas, the NBR and NDVI values of the consid-160 ered indicators after 10-15 years of restoration were 30-40% closer to the background 161 values compared to severely disturbed areas. However, even after 15 years a one-way 162 ANOVA test showed significant differences between disturbed and undisturbed areas 163 for both NBR (p < 0.005) and NDVI (p < 0.01) while before fires the differences between 164these areas were not significant. At the same time in case of NDVI the differences be-165 tween disturbed and undisturbed areas were less than 1 σ_b (Z < 1) for both severely and 166 moderately disturbed areas after 11 years post fire. For NBR event after 16 years after fire 167 the difference was higher than 1 σ_{b} (Z > 1). 168

4. Conclusions

Using the satellite data, an assessment of the fire-disturbed forest areas and post-fire 171 dynamics was performed for the southern regions of central Siberia. A decreasing trend 172 in forest burned area ($R^2 = 0.29$; p < 0.05) was observed in the region with mean burned 173 area of 250.2±153.7 thousand ha. The highest disturbance rates of 11.4% and 7.2% were 174 observed in mixed and larch-dominant forests. At the same time, the highest proportions 175 of severely burned areas were observed in the larch-dominant (20.8%) and DNC forests 176 (19.1%) while in the pine-dominant stands and mixed forests these were about two times 177 lower. 178

Fire frequency versus FRP was well-fitted by a power law ($R^2 = 0.92$, p < 0.05). The 179 highest FRP values (46.6±42.7 MW/km² and 42.8±44.1 MW/km²) were observed for the 180 larch-dominant and DNC forest types that is 25–30% than for pine-dominant and deciduous stands. While fire pixels within were characterized by significant dNBR variations the relationship between mean FRP and dNBR values was fitted using a logarithmic 183 law ($R^2 = 0.46$; p < 0.05). Fires resulted in severely burned areas generally had 72–116% 184 higher mean FRP value comparing to moderately burned areas. 185

The results indicated that severely disturbed areas need longer period of recovery. 186 For instance, after 10-15 years of recovery NBR and NDVI for severely disturbed areas 187 were 30–40% higher than for moderately disturbed areas. The differences between disturbed and background areas were still more than one standard deviation for NBR after 189 16 years of recovery, while for NDVI this difference becomes less than one standard deviation after 11 post fire. However, it takes more than 15 years for both indices to fully 191

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	recover after fire since even after 15 years since fire a one-way ANOVA test shows sig- nificant differences between disturbed and undisturbed areas.	192 193 194	
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	(<u>https://ladsweb.modaps.eosdis.nasa.gov</u>) (accessed on 22 September 2022).	201	
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