



Proceeding Paper

Measurement of the concentration of mercury for the automotive shredded residues using the direct mercury analyser†

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Abstract: Mercury content is among parameters that characterize quality of Solid Recovered Fuels (SRF), which constitute alternative solid fuels of increasingly interest. In this study, a direct mercury analyser is utilized in the analyses of SRF samples originating from Automotive Shredded Residues (ASR). Two SRF and one liquid reference materials are measured for testing the accuracy and sensitivity of the instrument. The mean values of the two SRF and one liquid reference materials (RM) were 0.171, 0.324 mg/kg and 0.141 mg/l. These values were compared with the reference ones: 0.168, 0.382 mg/kg and 0.137 mg/l. Mercury content is detected in analysis samples of four different particle sizes (8mm, 2mm, 1mm, 0.5mm and 0.25mm) for each SRF sample in order to demonstrate the direct mercury analyser operation and to compare it as an alternative to mercury determination via Atomic Absorption Spectrometry. The measurements showed that grinding down to 1 mm looks sufficient for most SRF cases, while grinding down to 2 mm might be enough for few cases. As an overall conclusion, the direct mercury analyser can be regarded as an efficient laboratory tool, offering a robust alternative to Atomic Absorption Spectrometry procedure, especially in terms of accuracy, speed, safety, and cost-effectiveness.

Keywords: Mercury; SRF; particle size; sample fraction; ASR; waste to energy

1. Introduction

Solid Recovered Fuels (SRF) refers to quality solid fuels derived from non-recyclable waste materials. They are produced from the non-biological dry fraction of waste streams, which are mainly those materials that cannot be recycled in an economically viable way. Automotive Shredded Residues (ASR) may be considered as Solid Recovered Fuels (SRF) provided that all hazardous materials have been removed, since the main requirement for solids to be considered as SRF is not to contain hazardous substances, according to ISO 21640 [1]. Mercury may be present in exhaust gases due to its high vapor tension. It has been identified as one of the world's most important environmental pollutants [2]. Recovering materials and energy from waste remains among European policies since they promote sustainable growth and circular economy [3]. ASR result from the shredding of End-of-Life Vehicles (ELVs) and is a by-product of the vehicle recycling process. ASR material is obtained having previously removed many recyclable vehicle parts, such as bumpers, airbags, batteries, fuel tanks, tires and seats. ASR is a very heterogeneous material consisting of plastics, elastomers, glass, wood, paper, leather, textiles, sand and metals [4].

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Copyright: © 2023 by the authors. Submitted for possible open access publication under the terms and conditions of the Creative Commons Attribution (CC BY) license (https://creativecommons.org/license s/by/4.0/). Vigano et al [5], reports that the range of mercury in ASR is between 0 to 0,5 mg/ kg on dry basis.

The analytical challenges presented by measuring mercury using techniques like cold vapor atomic absorption / atomic fluorescence (CVAA/AF), or Inductively Coupled Plasma- Optical Emission or Mass Spectrometry (ICP/OES-MS), are well recognized. The inherent problem lies in the fact that all these techniques are solution based, which means that if the sample is not a liquid it has to be digested before it is introduced to the instrument. Moreover, when heterogeneous samples are analysed, containing difficult to digest particles, it is becoming necessary to grind them down to smaller particles sizes than the suggested by the standard methods ones. Such size reduction is not an easy task when it comes to mixtures of plastic, elastomers, fibers and leather, such as SRF samples. Thus, the risk of less representative analysis samples becomes significant. Therefore, the utilisation of equipment requiring less elaborated sample pretreatment is a good option, provided it presents acceptable accuracy of results

A Direct Mercury Analyzer (DMA) is an instrument designed to measure the concentration of mercury, specifically in solid and liquid samples without the need for complicated sample preparation. The technique offers a rapid and simple method for the determination of mercury in alternative solid fuels [6]. Taking into consideration that mercury is a toxic element various methods have been developed to detect and quantify mercury in different samples. One of the classic and widely used methods for mercury determination is cold vapor atomic absorption spectrometry (CVAAS).

This study aims to prove the potential upgrade of the lab by highlighting the benefits of using the direct mercury analyzer instead of the traditional methods for mercury determination.

2. Materials and Methods

2.1. Direct Mercury Analyser: A Technique Overview

Direct mercury analyser (DMA) is an instrument designed to measure the concentration of total mercury directly in solid, liquid and even gas samples without the need for complicated sample preparation.

DMA requires minimal sample preparation since the samples can be introduced into the instrument directly and offers a rapid, within minutes, analysis. It reduces the use of toxic chemicals (such as HF solution for the digestion) and their potential wastes and also reduces the risk of contamination due to direct analysis and absence of wet chemistry [6].

In contrast, established techniques in elemental analyses, such as CVAAS and ICP, are more time-consuming due to the need for sample preparation (digestion of solid samples with toxic and hazardous solutions). Furthermore, the multiple steps following this technique can introduce sources of final result error, which may prove to be difficult to identify and eliminate. On the other hand, such techniques offer multi-element determination and have proved their accuracy [7].

2.2. Sample Preparation

The SRF samples were introduced in lab and they were dried in oven at 105°C for 24 hours. After drying the samples were grinded in pulverized mill with five different sieves (8mm, 2mm, 1mm, 0.5mm and 0.25mm). Each grinded sample was weighted and moved into the direct mercury analyser.

2.3. Testing DMA

In order to determine the efficiency of the newly acquired DMA in our lab, two sets of tests were done; one to identify its sufficient accuracy and the other to specify the required particle size of solid samples. SRF samples were used in both sets, bearing in mind that their results are probably valid for other solid fuels that are more homogeneous, like solid biofuels and fossil fuels.



Figure 1. A schematic of the DMA-80 evo direct mercury analyser [6].

Two SRF reference materials (RMs) were measured; both RMs are samples of interlaboratory test schemes. Their mean values were compared to the reference ones, taking into account the calculated uncertainty by the organiser of the schemes. Moreover, one liquid reference material was also measured using the same calibration and procedure.

The second set consists of testing three SRF samples originating from ASR. Each sample was tested after been grinded and sieved to 8 mm, 2 mm, 1 mm, 0.5 mm and 0.25 mm The 8 mm sub-sample is the raw sample in many cases. For each sample, the mercury concentration of the 0.25 mm particle size sub-sample is compared to the concentrations of the larger particle size sub-samples in order to determine up to what particle size one can accept DMA results as statistically indifferent. Compared sub-samples are considered different materials, having altered fuel properties. Hypothesis check is performed assuming that all samples are representative of populations (sources) where mercury concentration follows a normal distribution, while their variances are unknown and different. The null-hypothesis is that the populations of the compared sub-samples have the same mean mercury concentration. The alternative is that they do not. Therefore, the quantity td follows the t-distribution and is compared to the theoretical value of 2-side t-distribution with 95% probability and for v degrees of freedom [8]:

$$d = (m_1 - m_2) / s_d$$
 (1)

where m₁, m₂ are the mean values of the compared sub-samples and s_d is the typical error of the difference of the means:

$$s_{d} = \sqrt{(s_{1}^{2} / n_{1} + s_{2}^{2} / n_{2})}$$
(2)

where s₁, s₂ are the standard deviations of the compared sub-samples and n₁, n₂ are their measurements.

The degrees of freedom, v, required for the theoretical value of the t-distribution is calculated as:

$$v = 1 / (u^2 / (n_1 - 1) + (1 - u)^2 / (n_2 - 1))$$
(3)

where u is defined as

$$u = (s_{1^{2}} / n_{1}) / (s_{1^{2}} / n_{1} + s_{2^{2}} / n_{2}), \text{ for } s_{1^{2}} > s_{2^{2}}$$
(4)

Results and Discussion

The measurements of the SRF reference materials and of the liquid RM are presented in Table 1, followed by their reference values and the respective uncertainties. The specific names of the RM's are: SRF analytic proficiency test 2022 organised by Delta Coal Control GmbH (reference concentration: 0.1675mg/kg), SRF Schema 5121 proficiency test 2021 organised by Greek General Chemical State Laboratory (reference concentration: 0.382mg/kg) and waste water Schema 2412 proficiency test 2023 organised also by the Greek General Chemical State Laboratory (reference concentration 0.137mg/l). The DMA presents acceptable accuracy since the difference of the measured means and the reference values is smaller than the reported uncertainties.

Table 1. Mercury concentrations of reference material.

Quantity	SRF RM1 (mg/kg)	SRF RM2 (mg/kg)	Liquid RM (mg/l)
	0.164	0.311	0.144
	0.153	0.327	0.142
Measurements	0.175	0.306	0.135
Measurements	0.191	0.353	0.138
	0.161	0.306	0.145
	0.183	0.340	0.143
mean	0.171	0.324	0.141
ref mean	0.168	0.382	0.137
ref U	0.034	0.121	0.012
difference of means	0.004	-0.058	0.004

The measurements of the SRF samples that originate from ASR are presented in Tables 2 – 4. In each table, values of 5 sub-samples that differ in particle size are presented, their mean values and standard deviations, as well as the results of above defined quantities, td, sd and v (utilising Eq. 1 – 4). Each sub-sample is compared to one with nominal particle size of 0.25 mm. The theoretical value of the t-distribution, t_{th}, may be acquired from statistical tables. Comparing the absolute value of t_d with the 2-sided t-distribution value, we conclude with 95% probability that we must accept the null hypothesis for the 0.5 mm and 1 mm sub-samples, meaning that grinding down to 1 mm looks sufficient for most SRF cases, while grinding down to 2 mm might be enough for few cases, as illustrated in Table 2 (when sample heterogeneity can be regarded as low).

Table 2. Mercury concentrations of SRF sample 1 originating from ASR.

Quantity	d _p 8 mm (mg/kg)	d _p 2 mm (mg/kg)	d _p 1 mm (mg/kg)	d _p 0.5 mm (mg/kg)	d _p 0.25mm (mg/kg)
Measurements	0.429	0.534	0.535	0.526	0.539
	0.355	0.531	0.527	0.529	0.529
	0.393	0.498	0.526	0.541	0.53
	0.533	0.544	0.533	0.541	0.532
	0.45	0.481	0.535	0.542	0.539
	0.442	0.52	0.527	0.533	0.529
	0.447	0.528	0.532	0.539	0.536
	0.491	0.496	0.527	0.543	0.527
mean	0.443	0.517	0.53	0.537	0.533
std.dev	0.055	0.022	0.004	0.006	0.005
Sd	0.019	0.008	0.002	0.003	-
td	-4.652	-2.017	-1.084	1.553	-
ν	7	8	14	13	-
tth	2.364	2.306	2.145	2.160	-

Quantity	d _p 8 mm (mg/kg)	d _P 2 mm (mg/kg)	d _p 1 mm (mg/kg)	d _P 0.50 mm (mg/kg)	d _P 0.25 mm (mg/kg)
Measurements	0.337	0.386	0.418	0.431	0.426
	0.374	0.426	0.4	0.424	0.435
	0.369	0.398	0.421	0.433	0.426
	0.37	0.408	0.425	0.438	0.433
	0.347	0.4	0.409	0.424	0.427
	0.356	0.409	0.422	0.431	0.432
	0.365	0.417	0.434	0.438	0.437
	0.329	0.422	0.443	0.424	0.432
mean	0.356	0.408	0.421	0.43	0.431
std.dev	0.016	0.013	0.013	0.006	0.004
Sd	0.006	0.005	0.005	0.003	-
ta	-12.450	-4.517	-1.894	0.236	_
ν	8	8	8	13	_
tth	2.306	2.306	2.306	2.160	_

Table 3. Mercury concentrations of SRF sample 2 originating from ASR.

Table 4. Mercury concentrations of SRF sample 3 originating from ASR.

Quantity	d _p 8 mm (mg/kg)	d _p 2 mm (mg/kg)	d _p 1 mm (mg/kg)	d _p 0.50 mm (mg/kg)	d _p 0.25 mm (mg/kg)
Measurements	0.339	0.458	0.543	0.534	0.533
	0.411	0.456	0.542	0.54	0.543
	0.322	0.491	0.523	0.539	0.533
	0.405	0.508	0.511	0.539	0.527
	0.372	0.505	0.519	0.54	0.54
	0.364	0.496	0.512	0.528	0.528
	0.37	0.503	0.518	0.537	0.537
	0.407	0.496	0.548	0.529	0.544
mean	0.374	0.489	0.527	0.536	0.536
std.dev	0.033	0.021	0.015	0.005	0.007
Sd	0.012	0.008	0.006	0.003	-
ta	-13.788	-6.097	-1.460	0.110	-
ν	8	8	10	13	_
tth	2.306	2.306	2.228	2.160	_

The Direct Mercury Analyser (DMA) is an instrument that measures the concentration of total mercury in solid and liquid samples. Total mercury is trapped by an amalgamator as elemental mercury and subsequently measured by atomic absorption spectrometry, thus measurements may consider matrix independent and instrument calibration is sufficient for different sample materials.

Conclusion

As presented in this work, DMA requires minimal sample preparation compared to established techniques in elemental analyses while offering acceptable accuracy in the case of solid fuels. Analysis is complete within minutes and it does not require the use of toxic chemicals, eliminating their potential wastes, and reducing the analysis cost. The DMA can be regarded as an efficient laboratory tool, offering a robust alternative to Atomic Absorption Spectrometry and Induced Coupling Plasma Spectrometry procedures, especially in terms of accuracy, speed, safety, and cost-effectiveness.

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